CO₂ TO NATURAL GAS (CH₄) CONVERSION IN METHANOGEN-MINERAL SYSTEMS: THE ROLE OF MINERAL TYPE

by

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1. RESEARCH MOTIVATION AND PREMISE

Since the mid-19th century, atmospheric carbon dioxide (CO₂) levels have increased from less than 300 ppm to more than 400 ppm.¹ With more than 3.2 gigatons of CO₂ released annually, contributions from anthropogenic fossil fuel burning continues to be the primary source of excessive CO₂ inputs into the atmosphere.² As methods for discovery and recovery of oil and gas continue to be improved, fossil fuels are likely to remain a cheap energy source relative to current renewable energy strategies. The use of fossil fuels will therefore continue be a big part of our energy strategy into the foreseeable future and consequently will continue to negatively impact atmospheric CO₂ levels. Given current infeasibility in reducing atmospheric CO₂ levels via substantial reductions in fossil fuel use, mitigation strategies will need to include capacity to store large amounts of carbon dioxide from major CO₂ emitters.

In recent years, the capture and storage of CO₂ from the largest point source emitters in deep geological formations has emerged as the most promising mitigation strategy. This strategy, often referred to as geological CO₂ sequestration (GCS) or Carbon Capture Storage (CCS), has several advantages: 1) the technology needed is available and mature (the energy industry has been storing (and extracting) gases in (from) geological formations for decades, 2) the storage capacity needed is available (storage capacity in known suitable geologic formations within the conterminous US can hold all the CO₂ projected to be produced from fossil fuels over the next 2 centuries³, 3) environmental impacts are fairly well understood, can be assessed and minimized^{4,5} and 4) there are significant value-added potential both economically and scientifically.

The reality is that GCS is expensive and current government incentives in most instances are inadequate to encourage large-scale adaptation. This has led to calls from key

stakeholders (including the US Department of Energy) for the focus of CCS to be shifted from mere CO₂ capture and storage towards CO₂ utilization as an input/feedstock for valueadded processes. This addition of the "utilization" component to CCS (now CCUS) has seen successful efforts being made in the use of CO₂ in energy production efforts such as: enhanced oil recovery (CO₂-EOR) and enhanced coal-bed methane recovery (ECBM). Unmineable coal seams and oil/gas fields that were once considered non-economical have seen increases in productivity of 30-50 and 30-65 percent, respectively. 6-8 Although very positive for the overall value-added outlook of CCUS, CO₂-EOR and ECBM are limited by the fact that most of the available storage capacity for CO₂ is located in geological formations that contain no oil or coal-bed methane. Instead, the vast majority of CO₂ storage capacity is located in deep-saline aquifers with waters that are deemed too saline for domestic use and/or too uneconomical to pump. 9, 10 One possibility for adding value to CCUS in these deep saline aquifers is the enhanced conversion of CO₂ to methane (CH₄; the main component in natural gas) via autotrophic methanogenesis; with the subsequent utilization of the natural gas produced for energy generation.

The ubiquity of deep-saline aquifers and the widespread occurrence of microbial-mediated autotrophic methanogenesis in deep geologic environments^{11, 12} mean that the probability for locating enhanced CO₂-to-natural gas projects close to major point-source, CO₂ emitters (e.g. fossil-fuel-fired power plants and factories) are great. So too is the potential for incorporating the produced natural gas into a closed-loop hybrid power generation system; with CO₂ from the point source being pumped into the aquifer at one end and the natural gas produced (from enhanced CO₂-to-natural gas conversion) being used in part or full to power the plant or factory. Another associated benefit of enhanced CO₂-to-

natural gas conversion is the enhanced mineral trapping of CO₂, as a structural component, in carbonate minerals. Harvey *et al.*¹³ recently showed that the formation of iron-carbonate was enhanced in the presence of autotrophic methanogenesis but was absent from treatments without methanogenesis. The formation of carbonates from CO₂ is important because it results in the permanent sequestration of CO₂ in stable mineral deposits.

Despite the framework of methanogenesis-facilitated CCUS being plausible, the fundamental science needed to make this work is still in its infancy. For example, subsurface storage sites exhibit a great deal of heterogeneity with respect to mineral composition, pressure, temperature, and biological activity.² However, there is very little current research that integrates these factors. This research effort is intended to contribute to the fundamental science components needed to support a framework for enhanced CO₂-to-natural gas conversion as a value-added component in CCUS. Specific objectives will be: 1) to assess the effect of mineral electrical conductivity on the kinetics of CO₂-natural gas conversion; 2) to determine differences in rock/mineral interactions based on morphological characteristics (flagellated/non-flagellated) in order to better understand possible electron transfer mechanisms.

2. LITERATURE REVIEW

One limiting factor of synthetic conversion of CO₂ to CH₄ (autotrophic methanogenesis) is overcoming the very negative redox potential (-240 mV) of the reaction which requires catalysts and high energy input to produce substantial amounts of CH₄.¹⁴ Methanogenic archaea and bacteria are capable of overcoming this low potential by utilizing enzymes to facilitate CO₂ to CH₄ conversion in naturally-occurring reduced environments (subsurface geologic sites, anaerobic marshes, etc.) In addition to naturally-occurring

enzymes, several metals and metal oxides have been shown to increase rates of conversion by possibly acting as reducing-agents for these reactions in laboratory experiments.^{15, 16}

Several studies in the 1980's and 90's focused on the impacts of metal pollution on methanogenesis in marine environments. Capone et al. 14 studied the effects of a large suite of alkaline earth and transition metals on methanogenesis in homogenized salt marsh sediments. Most metals tested were found to enhance methanogenesis (after an initial lag time) with molybdenum (Na₂MoO₄) showing the greatest increase in maximum CH₄ production (23-fold increase in CH₄) compared to controls containing no metals. The increase in methanogenesis observed in these experiments was attributed to metal-related inhibition of SO₄²-respiring bacteria found in these syntrophic systems. ¹⁵ Lorowitz et al. ¹⁵ expanded on this research and tested the response of a single autotrophic species (Methanobacterium thermoautotrophicum) to a wide suite of elemental metals. This study showed that many metals could act as electron donors for this species significantly enhancing CH₄ production; for example, magnesium treatments had 954 µmol, iron treatments had 415 μmol, and zinc treatments 85 μmol compared with 0.76 μmol for control reactors. This study was important in showing that at least one species of autotrophic methanogen can utilize free electrons created by metals via a process known as cathodic depolarization. 16 It has been proposed that elemental metals can be oxidized in these aqueous environments which releases hydrogen that can be used in methanogenesis.

Methanococcus maripaludis is a hydrogeneotrophic archaea which thrives in strictly anaerobic environments such as deoxygenated swamps, subsurface pore space, and hydrocarbon reservoirs.¹⁷ In these oxygen-poor environments, *M. maripaludis* has the ability to utilize seven different hydrogenases to convert H₂ gas into free electrons, which are then

used to reduce CO_2 to CH_4 .¹⁸ The ability of this species to utilize simple carbon (CO_2 and formate) and nitrogen (N_2) substrates has spurred widespread interest for its potential use in carbon sequestration efforts. In addition, the optimal growth temperature is lower than many methanogenic species (37°C) and its growth rate (Population doubling time = 2 hrs.) is one of the fastest amongst methanogens making it a widely used model species.

M. maripaludis is small (0.9-1.3 μm) with a coccus structure which has two types of surface appendages: archaeal flagella (archaella) and pili. ¹⁹ The archaella have been shown to aid in motility in addition to surface attachment. Although pili appendages have not been as extensively studied in *M. maripaludis*, recent evidence suggests that both appendages are necessary for strong attachment to inorganic surfaces. ²⁰ Methanogenesis is the primary metabolism for the species as well as the only means of energy production for growth leading to competition for carbon resources. ¹⁸ It occurs via two mechanisms: disproportionation of formate or reduction of CO₂ with an electron donor such as hydrogen, formate, or electricity which can be shown in the simplified equations below: ^{18, 21}

$$CO_2 + 4H_2 \rightarrow CH_4 + 2H_2O$$

(CO_2 Reduction)

$$4HCOOH \rightarrow 3CO_2 + CH_4 + 2H_2O$$
 (Formate Disproportionation)

Although it has been studied for decades, the primary focus has been on genetic sequencing and alteration as well as its interactions with syntrophic organisms which aid in methanogenesis by means of extracellular electron transfer (EET)²². The studies that have incorporated metals and minerals with methanogens were primarily focused on environmental impacts regarding stimulatory effects on GHG's. However, recent research has shifted focus to the usefulness of methanogenic archaea such as *M. maripaludis* as

biological tools in mitigation efforts against climate change and specifically for CCUS. As mentioned earlier, methanogens have been shown to facilitate in the precipitation of iron-carbonate which could be useful in converting and storing large amounts of CO₂ without significant surface impact on the environment.¹³ Additionally, as populations grow wastewater treatment will continue to utilize methanogens for energy generation, which will be crucial to maintaining healthy urban societies in the future.^{23, 24} Finally, if natural gas prices were to increase or if the release of CO₂ were to become prohibitively expensive, methanogenic recovery of CH₄ from carbon sequestration may prove profitable in the future.²⁵

Recent evidence has shown that differences in mineral conductivity affects kinetic rates of conversion in soil settings. Kato *et al.*²⁶ and Zhou *et al.*²⁷ showed that methanogenesis in soils can be significantly influenced by the presence of semi-conductive iron oxide minerals. In these experiments, iron oxide minerals [ferrihydrite (5Fe₂O₃·9H₂O) , hematite (Fe₂O₃) , and magnetite (Fe₃O₄)] were added to rice paddy soils in the presence of a methanogens from the genus *Methanosarcina* as well as synotrophic *Geobacter* species.

Both experiments saw higher rates of methanogenesis in treatments containing minerals with higher Electrical Conductivity (*EC*). Zhou *et al.* showed increases in CH₄ production from 82.8 µmol in non-mineral controls to 172.9 µmol and 299.0 µmol in the presence of hematite ($EC = 3.45 \times 10^{-3} \, \text{mS m}^{-1}$) and magnetite ($EC = 4.09 \times 10^{-3} \, \text{mS m}^{-1}$), respectively. In contrast, there was significantly less CH₄ produced in the presence ferrihydrite ($EC = 2.69 \times 10^{-3} \, \text{mS m}^{-1}$) at 19.1 µmol CH₄. In addition to soil settings, metal amendments have also been shown to increase methanogenesis in anaerobic wastewater digesters which use

al. (2014) showed increases in CH₄ production of 43.5% when digesters were amended with zero-valent iron. Elemental manganese [Mn(0)] has also been shown to increase both rates and overall production of CH₄ in wastewater digesters. In 2015, Qiao *et al.* showed a rate increase in syntrophic aceticlastic methanogenesis of 3.5-fold and an overall production increase of 2.7-fold with respect to controls for amended treatments. One important finding from this study is that increases in production/activity were not directly linked to increases in concentrations of metals. (4 g/L > 8 g/L > 2 g/L). As mentioned before, it has been proposed that methanogens can use these metals/minerals as conduits for shuttling electrons during extra-cellular interspecies interactions during syntrophic methanogenesis. ^{26, 28, 29} In addition, autotrophic methanogenesis enhancement with elemental metals via cathodic depolarization has been proposed for decades based on the indirect evidence for this mechanism. ³⁰

There are no current studies, or at least to our knowledge, on intra-species shuttling of electrons or cathodic depolarization in methanogenesis in *M. maripaludis*. A large part of this may be attributed to the fact that, up until very recently, much of the research on methanogens (and particularly *M. maripaludis*) has focused on genetic sequencing and interspecies interactions. It can be reasonably inferred that if minerals have such drastic effects on methanogenic processes in marshes, soils, and wastewater, they must have some effect on subsurface methanogenesis where microbial communities have much greater contact with mineral surfaces. Since subsurface storage sites vary greatly in mineral composition (i.e. mineral conductivity), it is important to begin understanding the relationships between microbial communities and their surrounding environments. Although recent studies have introduced the relationships between minerals and syntrophic

methanogenic systems, few have focused on mineral/methanogen interactions in the absence of syntrophic organisms. Doing this, will determine whether methanogens themselves can use electrically conductive mineral surfaces to grow and/or increase methane production.

Although CO₂ methanogenesis occurs mostly via indirect hydrogenase-dependent mechanisms, recent evidence suggests that an alternative hydrogenase-independent pathway exists which may directly uptake electrons. A 2014 study by Lohner et al. reported methanogenesis in reactors containing M. maripaludis with genes omitted for hydrogenase production. In this study, hydrogenase-deleted mutants were able to draw electrons directly from an electrode at a rate of 1/10 of the rates observed in wild-type reactors ²¹. While this pathway is not well understood, it is possible that microbial communities are using this mechanism to shuttle electrons across conductive mineral surfaces in the absence of electron donors such as hydrogen gas. Jarrell et al. reported that both pili and archaealla are needed for attachment to mineral surfaces, but the range of uses for these appendages remains uncertain.²⁰ By deleting genes necessary for forming these appendages, they were able to show that wild-type Mm900 was able to attach to different inorganic surfaces (gold, nickel, molybdenum, glass) whereas mutants without pili and/or achaella did not attach readily to any surface. Since it is known that these appendages are used for surface attachment and a direct mechanism for electron uptake has been proposed, it is conceivable that M. maripaludis and other methanogens may be able to use appendages to shuttle electrons across conductive mineral surfaces when electron donors such as H₂ are scarce possibly through a cathodic depolarization mechanism. If this is the case, future research could focus on engineering this attribute to increase methanogenesis without the input of hydrogen which is both expensive and dangerous to work with on large scales.

3. MATERIALS AND METHODS

3.1. Mineral Selection and Preparation

To capture the full range of electrical conductivity likely to be encountered in geologic formations, minerals were selected using the work of Palacky (1988).³¹ In this study, earth materials were plotted as a function of electrical conductivity (Figure 1). For example, at the lower end of mineral electrical conductivity (herein referred to as, EC_{min}) limestone chalk (CaCO₃) and quartz (SiO₂) were chosen with EC_{min} range of 0.01-1 mS m⁻¹ and 0.1-2 mS m⁻¹, respectively. Kaolinite clay (Al₂Si₂O₅(OH)₄) served as the medium conductivity mineral with EC_{min} of 10-200 mS m⁻¹. Galena (PbS), and pyrite (FeS₂) were chosen as the high conductivity minerals found in reduced environments with EC_{min} of (1 x 10^3 - 1 x 10^5 mS m⁻¹).

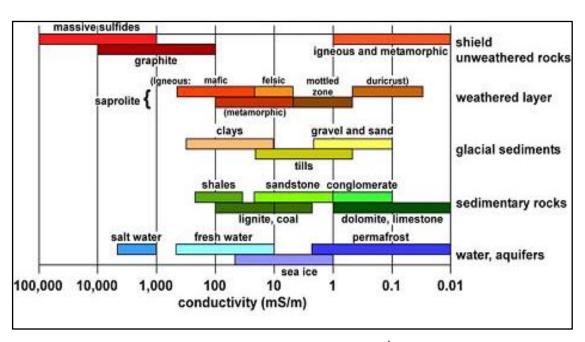


Figure 1. Electrical conductivity of rock units in mS m⁻¹ with massive sulfide rock units (galena and pyrite) located at the high end of the conductivity scale while limestone, quartz (sand), and kaolinite are located at the lower range of conductivity.³¹

Limestone and kaolinite mineral surfaces were cleaned with lint-free wipes; quartz, galena and pyrite samples were cleaned with distilled H_2O and dried with lint-free wipes. Samples were crushed separately using a mortar and pestle until fine enough to pass through a 500- μ m brass sieve. Galena and pyrite were subsequently cleaned with 0.5 M HCl and dried with N_2 gas in order to remove any oxides which may have been present on the mineral surface.

3.2. Sample Strain Cultivation

Two strains of M. maripaludis were used; Mm900 a wild-type/non-mutant strain and $\Delta flaB2$ a non-archaeallated mutant strain (which lacks genes necessary for archaella). Both strains were acquired as gifts from Dr. Kenneth Jarrell in the Department of Biomedical and Molecular Sciences at Queen's University. Both strains were cultured in 250-mL serum bottles (VWR International Item # 10170-770; Chemglass Life Sciences # 4217-03) based on modified methods provided by Drs. William Whitman and Feng Long of the Department of Microbiology at the University of Georgia. Details of these methods are provided in Appendix A. In brief, two 50-mL batches of each strain were cultivated in a stationary 37°C incubator prior to each set of experiments. The growth media was deoxygenated by sparging with N₂ gas and placed in an anaerobic chamber where 45 mL were dispersed into 250-mL serum bottles. Reactors were then sealed and crimped in the anaerobic chamber and autoclaved on the gravity cycle at 121°C for 25 minutes. After cooling, 1 mL of Na₂S·9H₂O solution (25 g/L) was added to each of the four reactors followed by inoculation of two reactors with 5 ml of Mm900, and two with 5 mL of Δ flaB2. Reactors were then pressurized with 80/20 v/v H₂/CO₂ gas to a pressure of 200 kPa, shaken for 30 seconds, and moved to the

37°C incubator room. Each reactor was re-pressurized and shaken every 24-36 hours during the incubation period.

3.3. Assessment of mineral effects on CO₂-to-CH₄ conversion by M. maripaludis

Two sets of experiments were conducted to assess mineral effects on CO₂-to-CH₄ conversion. Both sets of experiments were conducted in crimp-style anaerobic glass reactors in the presence and absence of H₂ (referred to as H₂-rich and H₂-limited conditions, respectively) and with mutant and non-mutant *M. maripaludis*. Each experimental set consisted of the following treatments: Minerals of varying conductivity with inoculum, non-mineral reactors with inoculum, gas blank reactors with N₂ gas instead of H₂ for hydrogen-rich experiments (and vice versa), and media blanks with no inoculum; with each treatment conducted in triplicates.

Experimental set 1 was focused on identifying general growth trends with respect to mineral conductivity in H_2 -rich and H_2 -limited environments and included the full complement of minerals: limestone, quartz, kaolinite, pyrite, and galena. These experiments were carried out in 28 mL glass "Balch" tube reactors (VWR International #89167-180; Chemglass Life Sciences #4209-10) for incubation periods of 1 and 14 days. For mineral treatments, 0.5 ± 0.01 g of each mineral were added to the reactors in triplicates. The reactors (including those for gas and media blanks) were then covered and transferred into an anaerobic chamber (VAC 101965 OMNI-LAB) along with 1 L of freshly prepared growth mediim. While in the anaerobic chamber, 4 mL of archaea growth medium was transferred to each reactor receiving *M. maripaludis* cultures and 5 mL to medium blanks which would receive no cultures. Reactors were then sealed and crimped with butyl rubber stoppers and aluminum seals. Stoppers and aluminum seals were acquired from VWR International (Item

30624-022 and 89167-1834) and Chemglass Life Sciences (item # 4209-14 and 4209-12). Crimped reactors were removed from the chamber, autoclaved (121°C on gravity setting for 25 minutes), cooled to room temperature, the top of each tube flame-sterilized and 0.1 mL of Na₂S·9H₂O solution injected into the reactors. Reactors were then shaken vigorously and allowed to sit at room temperature for a period of 18 hours.

After the 18h sitting period, mineral and non-mineral *M. maripaludis* treatments were inoculated with 1 mL of mutant or non-mutant strains. Reactors for H₂-rich treatments were then flushed for 1 minute, pressurized to 200 kPa with H₂/CO₂ gas mixture (certified 80% H₂ and 20% CO₂ mix) and incubated for 1 and 14 days at 37 °C. Reactors for H₂-limited treatments, were treated similarly except instead of the 80% H₂ /20% CO₂ gas mix, an 80% N₂ /20% CO₂ gas mix was used for flushing and pressurizing the reactor headspace. pH measurements were taken on each sampling, but were not used in data analysis (Appendix B).

Experimental set 2 was focused on assessing kinetics of R_{CH4} (CH₄:CH₄+CO₂) across minerals under H₂-rich and H₂-limited conditions. These experiments were similar in many aspects to experimental set 1 except reactors were 250-mL glass serum bottles (rather than 28-mL Balch tubes used in experimental set 1) and sampling of headspace CH₄ and CO₂ was taken on day 1, 2, 3, and 4 (not just day 1 and 14). Experiment 2 also include only limestone and pyrite - minerals from both extremes of the electrical conductivity spectrum. In order to maintain the same mineral-to-solution ratio (~1:10) as in experimental set 1 mineral mass and volumes for archaea growth media, inoculum and Na₂S·9H₂O were increased by a factor of 10. Mineral treatments in experimental set 2 contained 5g of mineral, 45 mL of growth media, 5 mL inoculum and 1 mL of Na₂S·9H₂O solution. As in experimental set 1, reactors

were flushed and pressurized to 200 kPa with 80/20 gas mixture of H_2/CO_2 (for H_2 -rich treatments) or N_2/CO_2 (for H_2 limited treatments) and incubated at $37^{\circ}C$.



Figure 2. Experimental Setup in Anaerobic Chamber where media was transferred to "Balch" tubes with varying treatments followed by crimping and sealing.

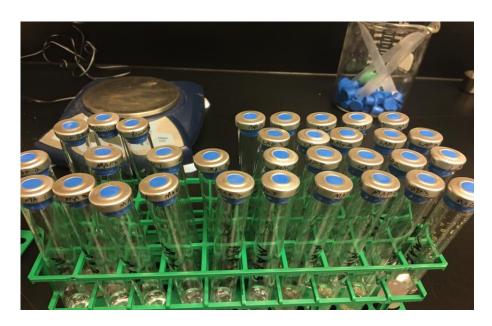


Figure 3. Sealed and crimped reactors prior to autoclaving, inoculation, and pressurizing with H_2/CO_2 or N_2/CO_2 gas.

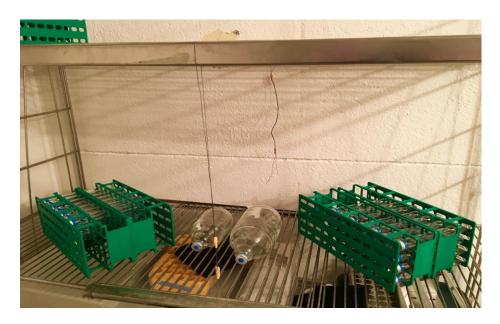


Figure 4. Incubation of reactors in a 37°C incubator room.

3.4 Headspace Gas Sampling and Analysis

All headspace gas measurements were made using gas chromatography. The gas chromatograph used was a "greenhouse" gas GC (SRI 8610C) from SRI instruments. Sample introduction into the instrument was facilitated via a 10-port gas sampling valve with 1-mL sampling loop. Gas separation was via a 15' HAYESEP-D column with N₂ as the carrier gas. For gas detection, the instrument was equipped with both an electron capture (ECD) and a flame ionization detector (FID). Operating conditions for the instrument during CO₂ and CH₄ analysis was 100 °C oven temperature and carrier flow was 30 psi nitrogen. Average elution time for CO₂ and CH₄ were 2.00 and 3.15 minutes respectively. Calibration curves for CO₂ and CH₄ were established daily using calibration gases containing 5 and 20% of CO₂ and CH₄. Calibration gases were purchased from GASCO Gas Company and were run in triplicates.

To introduce the gas (calibration or headspace gas) sample into the sample loop, a 5 – mL syringe was connected to the outlet side of 10-port valve - with the sample to be tested connected to the inlet side. For the calibration gases, connection was directly through a tube from the valve of the calibration gas canister to the inlet of 10-port valve (Figure 5a). For headspace gas sampling, the reactors were connected to the inlet side of the 10-port via a flame-sterilized syringe needle plunged through the top of the flame-sterilized crimp seal and into the headspace (Figure 5b). The piston of the 5-mL syringe on the outlet end was then pulled from the 0 to 3 mL position, allowing for filling of 1-mL sample loop. Once the sample loop was filled, the valve was activated to the "inject" mode resulting in injection of the sample onto the 15' HAYESEP-D column.



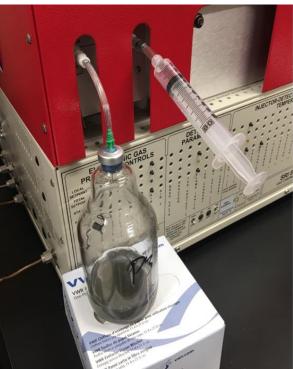


Figure 5. Gas chromatograph injection of a) gases drawn directly into GC with plastic syringe. b) reactor headspace drawn through vacutainer needle into the loop with syringe calibration.

Peaks for CO_2 and CH_4 on the resulting chromatographs were integrated and compared to those for calibration gases to determine CO_2 and CH_4 concentration, and the concentrations were used to calculate the relative CO_2 -to- CH_4 conversion ratio (R_{CH4}) for each treatment. Relative CO_2 -to- CH_4 conversion ratio was calculated as:

$$R_{CH4} = \frac{[CH_4]}{[CH_4] + [CO_2]}$$

where, $[CH_4]$ and $[CO_2]$ are the average concentrations of CO_2 and CH_4 respectively across triplicate reactors within a given experimental treatment. Overall concentrations of CH_4 and CO_2 were not used for these analyses, but are reported in Appendix C.

4. RESULTS

4.1. Mineral electrical conductivity (EC_{min}) effects on CO₂-to-CH₄ conversion ratio (R_{CH4}) by M. maripaludis

4.1.1 Insights on EC_{min}-R_{CH4} relationship from experiments in H₂-rich environments

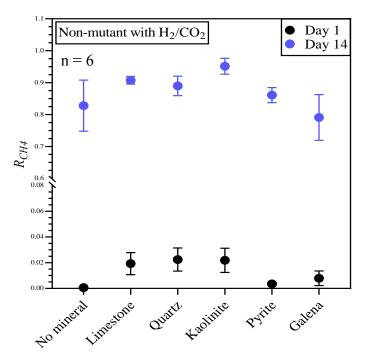


Figure 6. R_{CH4} for wild-type Mm900 across varying mineral type in H_2/CO_2 for 1 and 14 with days

Figure 6 shows the R_{CH4} values in the reactor headspace for days 1 and 14 from experimental set 1 under H₂-rich conditions. Minerals are arranged from left-to-right in order of increasing electrical conductivity. I.e. the average electrical conductivity for the minerals follow the order limestone <quartz< kaolinite (clay) < pyrite (sulfide) \le galena (sulfide). Figure 6 indicated that within the first day of incubation, there was no detectable quantity of CH₄ among the total carbon gas (CO₂+ CH₄) measured in the headspace of reactors containing no mineral and the high electrical sulfide minerals (galena or pyrite). This was consistent with an absence of CO₂-to-CH₄ conversion by Mm900 on day 1 under H₂-rich condition in these treatments. In contrast, 2-3% (R_{CH4} = 0.02 – 0.03) of the total carbon gas in the low conductivity minerals (quartz, limestone and kaolinite) was CH₄ - suggesting that at least in the early stages of the experiment, CO₂-to-CH₄ conversion by Mm900 under H₂-rich conditions was inversely related to mineral conductivity.

By day 14, R_{CH4} in the non-mineral reactors increased to 0.89 (from 0 on day 1) in the reactor headspace and was comparable to that in the quartz-, limestone- and kaolinite-containing reactors – with R_{CH4} of 0.89, 0.91 and 0.95, respectively (Figure 6). Galena- and pyrite-containing reactors had significantly lower R_{CH4} (0.79 and 0.86, respectively) – consistent with the qualitative trends observed on day 1. The fact that the qualitative trend in the mineral conductivity- R_{CH4} relationship from day 1 and day 14 was similar throughout the experiment (i.e. lower electrical conductivity of minerals favors higher R_{CH4}) suggested that, under H_2 -rich conditions, CO_2 -to- CH_4 conversion in the Mm900 strain was being enhanced in the presence of lower versus higher conductivity mineral in H_2 -rich environments. It is also worth noting that average concentrations of CH_4 produced by Mm900 in the headspace of H_2 -rich reactors were congruent with this trend. For example, on day 14 the reactors

containing limestone, quartz and kaolinite, averaged 4-7% CH₄ compared to 1-3% in pyriteand galena-bearing reactors.

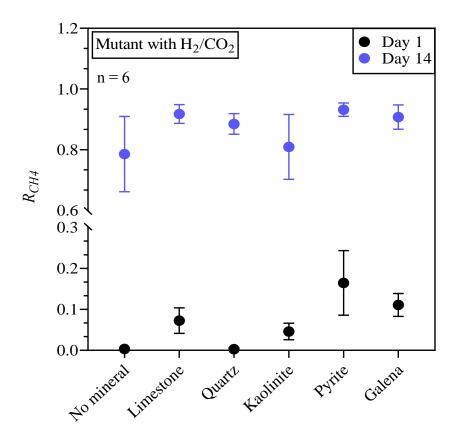


Figure 7. R_{CH4} for non-archaellated $\Delta flaB2$ with varying mineral type in H_2/CO_2 for day 1 and 14

Very different mineral- R_{CH4} relationships were apparent in H₂-rich conditions for the non-archaellated $\Delta flaB2$ than was observed for the archaellated Mm900 strain. In contrast to the inverse relationship observed between mineral conductivity and R_{CH4} for Mm900 (Figure 6), for $\Delta flaB2$, R_{CH4} on day 1 was generally lower ($R_{CH4} = 0$ -0.07) in reactors containing the low-conductivity minerals (limestone, quartz and kaolinite) than in those containing the high conductivity minerals (pyrite and galena; $R_{CH4} = 0.11$ -0.17). No clear trend between mineral conductivity and R_{CH4} was apparent in the day 14 data for $\Delta flaB2$ (Figure 7) suggesting that the positive mineral conductivity- R_{CH4} relationship observed on day 1 could (at least

qualitatively) be linked to enhancing effects of mineral conductivity on the early stages of CO_2 -to- CH_4 conversion by $\Delta flaB2$ under H_2 -rich conditions.

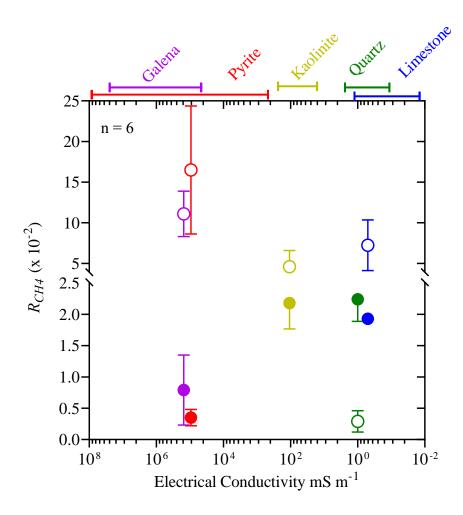


Figure 8. CO₂-to-CH₄ conversion ratio (R_{CH4}) by Mm900 (filled circles) and $\Delta flaB2$ (open circles) as a function of mineral electrical conductivity on day 1 in H₂-rich conditions.

Plotting R_{CH4} (for Mm900 and $\Delta flaB2$) as a function of median EC_{min} on the same log-scale as Figure 1 suggested that effects on R_{CH4} was linearly related to the log-transformed EC_{min} (Figure 8). That is, the R_{CH4} - EC_{min} relationship for Mm900 and $\Delta flaB2$ in H₂-rich conditions could be described by the general equation: $R_{CH4} = m \log(EC_{min}) + c$; where the slope, m, of relationship indicates the direction and magnitude of the EC_{min} effects while the

intercept, c, corresponds to the R_{CH4} value expected in the presence of a mineral with EC_{min} of unit conductivity units (in this case, 1 mS m⁻¹). Table 1 shows the R_{CH4} -log(EC_{min}) relationships for the minerals studied. For calcite, quartz and kaolinite the values used for median EC_{min} (0.5,1, and 100 mS m⁻¹, respectively) were obtained from ranges for limestone, sand and clay in Figure 1. Median EC_{min} values for galena (1.5 x 10⁵ mS m⁻¹) and pyrite (9.3 x 10⁴ mS m⁻¹), were calculated from the range of resistivity measurements presented by Pridmore and Shuey (1976) for these minerals. ³²

Table 1 shows the R_{CH4} -log (EC_{min}) equations for day 1 and day 14 CO₂-to-CH₄ conversion by Mm900 and $\Delta flaB2$ under H₂-rich conditions. These relationships confirmed the potential quantitative links between EC_{min} and the conversion of CO₂ to CH₄ by Mm900 and $\Delta flaB2$ under H₂-rich conditions. The day 1 values for Mm900 showed a statistically significant (p-value = 0.032) non-zero slope with the equation of R_{CH4} = -(2.99 x 10⁻³)(log(EC_{min})) + 0.022 (R² = 0.829). This indicates that under H₂-rich conditions Mm900 exhibited significantly different trends in presence of minerals. The R_{CH4} -log(EC_{min}) relationships for day 14 did not show significant non-zero slopes (and therefore was not discussed in this paper). The values for $\Delta flaB2$ in H₂-rich conditions did not show significance in either day of the experiment (1 or 14).

Table 1. Day 1 equations and fits for R_{CH4} -log (EC_{min}) relationships for Mm900 and $\Delta flaB2$ in H₂-rich and H₂-limited environments.

Equations for R_{CH4} -log (EC_{min}) Relation
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H ₂ -rich conditions (day 1)	$m \log(EC_{min}) + c$	<i>p-value</i> (<i>m</i> ≠ 0)	R ²
<u>Mm900</u>	$-2.99 \times 10^{-3} \log(EC_{min}) + 0.022$	0.032	0.829
<u>ΔflaB2</u>	$1.91 \times 10^{-2} * \log(EC_{min}) + 0.035$	0.092	0.666
H ₂ -limited conditions (day 14)			
<u>Mm900</u>	$1.60 \times 10^{-2} * \log(EC_{min}) + 0.917$	0.182	0.500
<u>ΔflaB2</u>	$5.47 \times 10^{-3} * \log(EC_{min}) + 0.038$	0.626	0.089

4.1.2. Insights on EC_{min} - R_{CH4} relationship from experiments in H_2 -limited environments

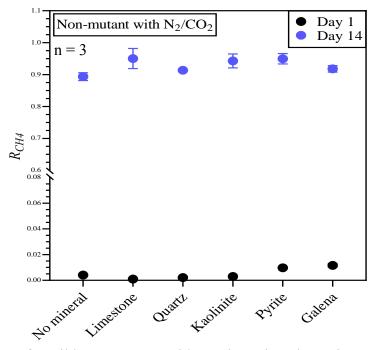


Figure 9. R_{CH4} for wild-type Mm900 with varying mineral type in N₂/CO₂ for days 1 and 14.

Figure 9 shows the R_{CH4} values for day 1 and 14 for Mm900 in H₂-limited environments. Under these conditions, no mineral and lower-conductivity reactors showed low R_{CH4} values ranging from 0.00-0.04. Galena and pyrite reactors produced significantly higher R_{CH4} ratios during this period with 0.11 and 0.10, respectively. High-conductivity minerals produced significantly higher R_{CH4} values during this time period compared with reactors with no minerals or minerals of lower conductivity. This was opposite of the day 1 trend seen in H₂-rich Mm900 reactors with a trend directly related to mineral conductivity (i.e. higher EC = higher R_{CH4}). By day 14, in non-mineral reactors CH₄ accounted for 0.89 of gaseous carbon which was comparable to mineral-containing reactors with quartz-, limestone-, kaolinite-, galena-, and pyrite-containing reactors having R_{CH4} values ranging from 0.91 to 0.95 with no trend relative to mineral conductivity. With the exception of quartz, the highest final CH₄ concentration under H₂-limited conditions with wild-type Mm900 was congruent with the trend in R_{CH4} . Reactors containing galena had day 14 CH₄ concentrations of (22.73%) followed by quartz (20.88%), pyrite (19.39%), kaolinite (18.30%), non-mineral (16.83%) and limestone (12.94%).

On day 1, the ratio of gaseous carbon as CH₄ for hydrogen-limited reactors containing mutant strain $\Delta flaB2$ and no minerals was 0.01 which was comparable to limestone (0.00) and kaolinite reactors (0.01) as seen in Figure 10. Quartz reactors produced a relatively high R_{CH4} value under these conditions with 0.13. In general, R_{CH4} was higher in reactors containing minerals with higher conductivity (R_{CH4} for galena and pyrite was 0.07 and 0.54, respectively) on day 1. By day 14, CH₄ represented 92-95% (R_{CH4} = 0.92-0.95) of all gaseous carbon in the headspace of the reactors showing no trend with mineral conductivity. Final overall CH₄ concentrations for non-archaellated $\Delta flaB2$ were highest in reactors

containing limestone (22.54%) followed by kaolinite (19.28%), non-mineral (18.90%), pyrite (15.34%), galena (14.13%), and quartz (9.23%).

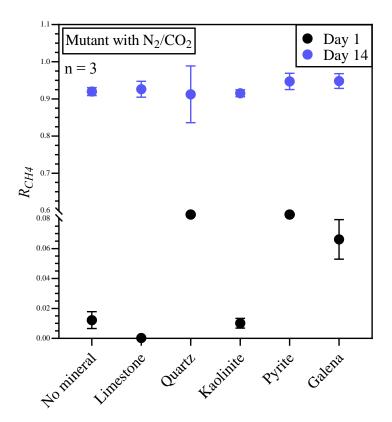


Figure 10. R_{CH4} for non-archaellated $\Delta flaB2$ with varying mineral type in N₂/CO₂ for days 1 and 14.

As with H₂-rich environments, R_{CH4} values (for Mm900 and $\Delta flaB2$) showed linear relationships when plotted as a function of median EC_{min} on a log scale (Figure 11). Table 2 shows R_{CH4} -log (EC_{min}) relationships and respective equations for day 1 CO₂-to-CH₄ conversion by Mm900 and $\Delta flaB2$ under H₂-limited conditions. Day 1 relationships for Mm900 showed a significant (p-value = 0.003) non-zero slope with the equation of R_{CH4} = $(1.80 \times 10^{-3})(\log(EC_{min})) + 0.001$ (R² = 0.961). This indicates that under H₂-limited conditions Mm900 exhibited significantly different trends in presence of minerals (which was also the case with Mm900 in H₂-rich conditions). Again, the R_{CH4} relationships for day 14

did not show significant non-zero slopes and therefore were discussed in this paper. The plots for $\Delta flaB2$ in H₂-limited conditions once again showed no significance in either day of the experiment (1 or 14).

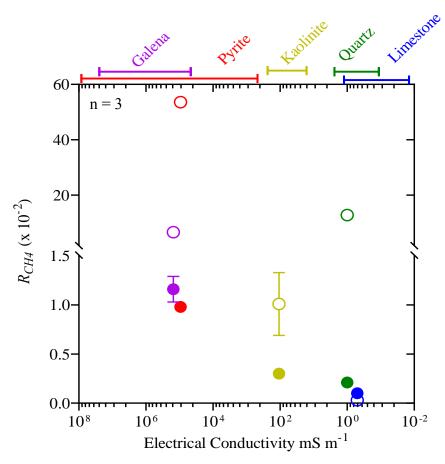


Figure 11. CO₂-to-CH₄ conversion ratio (R_{CH4}) by Mm900 (filled circles) and $\Delta flaB2$ (open circles) as a function of mineral electrical conductivity on day 1 in H₂-limited conditions.

Table 2. Day 14 equations and fits for R_{CH4} -log (EC_{min}) relationships for Mm900 and $\Delta flaB2$ in H₂-rich and H₂-limited environments.

Equations for R_{CH4} -log (EC_{min}) Relationships				
H ₂ -rich conditions (day 1)	$m \log(EC_{min}) + c$	<i>p-value</i> (<i>m</i> ≠ 0)	R ²	
<u>Mm900</u>	$1.80 \times 10^{-3} \log(EC_{min}) + 0.001$	0.003	0.961	
<u>∆flaB2</u>	$4.75 \times 10^{-2} * \log(EC_{min}) + 0.038$	0.323	0.317	
H ₂ -limited conditions (day 14)				
<u>Mm900</u>	$2.09 \times 10^{-4} * \log(EC_{min}) + 0.935$	0.182	0.500	
<u>ΔflaB2</u>	$5.65 \times 10^{-3} * \log(EC_{min}) + 0.917$	0.052	0.766	

4.1.3 Summary of general trend analysis for Experiment 1

Qualitative trends for experiment 1 revealed that in most (3/4) conditions reactors containing high-conductivity minerals outperformed non-mineral and low-conductivity reactors with respect to R_{CH4} values. However, only day 1 values for reactors containing Mm900 showed significant non-zero slope differentiation for these plots. Under H₂-rich conditions, CO₂-to-CH₄ conversion by Mm900 exhibited a trend inversely related to mineral conductivity with low-conductivity minerals (limestone, quartz, and kaolinite) enhancing methanogenesis in early-stages (day 1). In contrast, under H₂-limiting conditions this trend in R_{CH4} was reversed with high-conductivity minerals (pyrite and galena) enhancing early-stage methanogenesis. This reversal of trends suggests that in the presence of hydrogen, electrons may preferentially be used for growth (not measured in these experiments) instead

of maintenance. In contrast, the H₂-limited trend suggests that in the absence of hydrogen, conductive minerals may act as electron donors in the early-stages possible via cathodic depolarization as reported in elemental metals.³³

An increase in early-stage methanogenesis involving mutant strain $\Delta flaB2$ was seen in reactors containing galena and pyrite in both hydrogen-rich and hydrogen-limited reactors (as well as non-mutant H₂-limited reactors). This result would not be expected if archaella are involved in the direct uptake mechanism established by Lohner *et al.*¹⁵ or any indirect mechanism involving electron shuttling by *M. maripaludis.*²¹ However, non-zero slopes for R_{CH4} -log (EC_{min}) relationships in $\Delta flaB2$ reactors were not significant showing that qualitative trends seen in non-archaellated reactors were not significantly different than non-mineral controls.

4.2. Mineral effects on kinetics of CO₂-to-CH₄ conversion

Results from Experimental set 1 highlighted several important aspects of mineral effects on conversion of CO₂-to-CH₄ in H₂-rich and H₂-limited environments. For example, early-stage methanogenesis was found to be significantly impacted by mineral type in H₂-rich environments, but not in H₂-limited environments. In addition, these significant trends were opposing with respect to mineral conductivity with limestone reactors having higher R_{CH4} values than pyrite under H₂-rich conditions (and vice versa- under H₂-limited conditions). One observation of primary importance is the fact that any differences across mineral treatments were significant in day 1 (but not in day 14) - suggesting that mineral effects were most likely kinetically driven. Data from experimental set 2 were used to assess mineral effects on the kinetics of CO₂-to-CH₄ conversion by Mm900 and $\Delta flaB2$ in both H₂-rich and H₂-limited conditions. Experiments included non-mineral, limestone- and pyrite-

containing treatments. Limestone and pyrite were used to represent the lower and higher end of the mineral conductivity spectrum, respectively. The much larger volume of the 250-mL reactors (compared to 28 mL in Experimental set 1) used in these experiments allowed for 1) a significant reduction in the number of reactors needed and 2) multiple sampling of each headspace-thereby reducing intra-experiment variability associated with sampling.

4.2.1. Temporal trends and kinetics of CO₂-to-CH₄ conversion

Figure 12 shows the temporal evolution in R_{CH4} for Mm900 and $\Delta flaB2$ in non-mineral, limestone- and pyrite-containing reactors under both H₂-rich conditions and H₂-limited conditions. In all cases, R_{CH4} peaked and leveled off within 4 days of initiation - indicating that CO₂-to-CH₄ conversion was completed within 4 days. Peak R_{CH4} were always between 0.80 and 0.90, congruent with those observed in experimental set 1 and further confirmed that observed differences were dictated by differences in CO₂-to-CH₄ conversion kinetics. Time to peak R_{CH4} across treatments followed the general order limestone> non-mineral> pyrite. In contrast, trends in day-1 R_{CH4} followed the order pyrite > non-mineral> limestone consistent with the overall rate of CO₂-to-CH₄ conversion being highest in pyrite-containing and lowest in limestone-containing reactors.

A logistic (sigmoidal) model was used to estimate key kinetic parameters for CO_2 -to- CH_4 conversion mediated by Mm900 and $\Delta flaB2$ in the presence/absence of limestone/pyrite under H_2 -rich and H_2 -limited conditions. The model was fitting using Graphpad Prism© software and was of the form:

$$R_{CH4} = \frac{c}{1 + e^{-k(t - t_{1/2})}}$$

27

where R_{CH4} is the conversion ratio for CO₂ to CH₄; c is the maximum conversion captured by the model; k is the pseudo-first order rate constant in units of day⁻¹; t is the experimental time/length of incubation in units of days; $t_{1/2}$ is the time corresponding to half of R_{CH4} captured by the experiment/model.

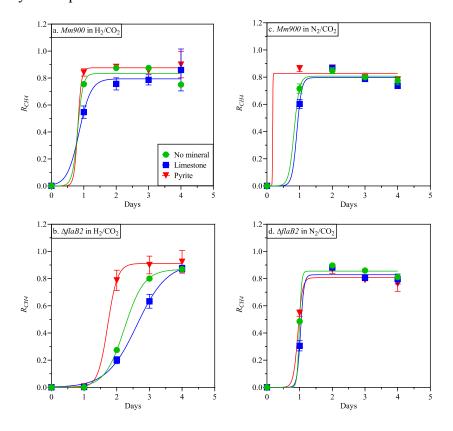


Figure 12. Kinetics of CO₂-to-CH₄ conversion (R_{CH4}) in H₂-rich and H₂-limited environments with wild-type Mm900 and non-archaeallated $\Delta flaB2$ with respect to non-mineral(green), limestone (blue) and pyrite (red).

Model fits to the R_{CH4} versus time data reflected good fits for all scenarios plotted (Figure 12). Values for co-efficient of determination, R^2 ranged between 0.966 and 0.999 indicating that the model captured 97-99% of variability in R_{CH4} with time (Table 2). Table 2 also summarizes the fitted parameters for the model. That is, c, k and $t_{1/2}$.

Table 3. Modeled parameters max conversion (c), rate constant (k), and half-life ($t_{1/2}$) with r^2 values. Standard error has been reported where applicable.

Modeled Parameters				
H ₂ -rich conditions	С	Rate constant, k	t _{1/2}	\mathbb{R}^2
<u>Mm900</u>				
No Mineral	0.834 ± 0.021	12.0	0.811	0.986
Limestone	0.795 ± 0.024	5.43	0.854 ± 0.138	0.966
Pyrite	0.876 ± 0.013	17.3	0.815	0.991
∆flaB2				
No Mineral	0.869 ± 0.008	3.32 ± 0.172	2.24 ± 0.018	0.999
Limestone	0.915 ± 0.026	2.17 ± 0.162	2.62 ± 0.049	0.995
Pyrite	0.912 ± 0.022	6.66	1.72 ± 0.181	0.988
H ₂ -limited conditions				
<u>Mm900</u>				
No Mineral	0.808 ± 0.012	12.3	0.832	0.991
Limestone	0.797 ± 0.017	13.6	0.917	0.98
Pyrite	0.829 ± 0.013	181	0.175	0.987
∆flaB2				
No Mineral	0.855 ± 0.011	28.5	0.991	0.993
Limestone	0.829 ± 0.016	18.7	1.03	0.989
Pyrite	0.807 ± 0.016	14.3	0.948	0.982

In H_2 -rich environments containing wild-type Mm900, the highest c value was found in reactors containing pyrite (0.876 ± 0.013) followed by reactors containing no minerals (0.834 ± 0.021) and limestone (79.5 ± 0.024) . In similar environments containing nonarchaellated $\Delta f laB2$, pyrite reactors produced slightly lower c values than limestone with 0.912 ± 0.022 and 0.915 ± 0.026 , respectively. Reactors with no minerals had the lowest c values under these conditions with 0.869 ± 0.010 . The trends in H₂-rich environments do not match across archaellated and non-archaellated treatments. In H₂-limited environments containing wild-type Mm900, the highest c value was found in reactors containing pyrite once again (0.829 \pm 0.013) which was slightly higher than both non-mineral (0.808 \pm 0.012) and limestone (0.797 \pm 0.017) reactors. H₂-limited reactors containing $\Delta flaB2$ and pyrite produced the lowest c values (0.807 \pm 0.016). Non-mineral reactors produced the highest values under these conditions (0.855 \pm 0.011) followed closely by limestone (0.829 \pm 0.016). In both environments, wild-type *M. maripaludis* produced the highest maximum conversion following a trend in reactor mineral conductivity; pyrite>non-mineral>limestone. Conversely, pyrite reactors containing $\Delta flaB2$ produced slightly lower c values than limestone reactors in both H₂-rich and H₂-limited environments.

The rate constant (k) for wild-type Mm900 in H₂-rich environments was highest in pyrite reactors (~17.3) and followed the same trend that was seen in the c values. Non-mineral reactors had the second highest k values (~12.0), followed by reactors with limestone (5.43). Unlike the opposing trends seen with c values (with respect to Mm900), k-value trends for $\Delta flaB2$ in H₂-rich environments were similar to those seen in the wild-type strain; pyrite>non-mineral>limestone. Reactors containing pyrite had the highest rate constant (6.66) followed by non-mineral (3.32 \pm 0.17) and limestone (2.17 \pm 0.16) reactors. In H₂-

limited environments, k values for wild-type M. maripaludis in reactors containing pyrite were unrealistic. However, non-mineral and limestone reactors exhibited trends opposite those seen in H₂-rich environments (Limestone>non-mineral). The rate constant for limestone reactors was ~13.56 which was slightly higher than non-mineral reactors (~12.25) under these conditions. k values for reactors with $\Delta flaB2$ were highest in reactors with no minerals (~28.53) followed by limestone and pyrite reactors at ~18.72 and ~14.27, respectively. No trend between rate constant and mineral conductivity could be determined under these conditions.

The half-life parameter $(t_{1/2})$ is slightly different than the other two parameters in that the lower value corresponds to the fastest time. This is the amount of time in days that it takes for 50% conversion of CO₂ to CH₄ in each reactor. The fastest half-life in H₂-rich reactors containing Mm900 was seen in the absence of minerals (~0.811 days) followed by reactors containing pyrite (~ 0.815 days) and limestone (0.854), respectively. In reactors containing $\Delta flaB2$, those containing pyrite had the fastest half-life (1.724 \pm 0.18 days) followed by non-mineral reactors (2.236 \pm 0.02 days) and limestone reactors (2.616 \pm 0.05). In H₂-limited conditions, half-life values for reactors containing Mm900 and pyrite were unrealistic and therefore were not plotted. Non-mineral reactors had the shortest half-life at ~0.832 days while limestone reactors were slightly slower during the experiment (~0.917 days). Under H₂-limited conditions, reactors containing $\Delta f laB2$ had the shortest half-life in the presence of pyrite (~0.948 days) followed closely by non-mineral (~0.991 days) and limestone (\sim 1.029 days) reactors. In all mineral reactors, $t_{1/2}$ was shortest in the presence of pyrite (with the exception of Mm900 H₂-limited) and lowest with limestone. However, nonmineral reactors had the shortest half-life in mutant reactors under both conditions.

CONCLUDING REMARKS

The majority (7/8) of reactors sampled in our experiment saw increases in early-stage methanogenesis in the presence of high-conductivity minerals. Inhibition of SO₄²—respiring bacteria (as seen in syntrophic communities) can be ruled out as the mechanism for increased methanogenesis because only *M. maripaludis* was present in each treatment.¹⁴ Although debated frequently, a mechanism involving cathodic depolarization seems the most likely to be causing increases in methanogenesis in the presence of high-conductivity minerals. ¹⁵ Most reactors showed increases in CO₂-to-CH₄ conversion in the presence of these minerals, but it is the H₂-limited treatments that are the most interesting. The fact that reactors saw similar trends with increasing conductivity in these environments suggests that enhancement is derived from the minerals irrespective of the presence of H₂ which fits with cathodic depolarization theory. Limestone, quartz, and kaolinite (low-, med-conductivity) are not made of redox active metal ions like those seen in pyrite and galena, so the trends seen through these experiments are plausible if cathodic depolarization is occurring. Future studies should focus on these mineral/methanogen interactions, and also the interactions between minerals and varying environments in the absence of methanogens. Although unlikely to occur in the highly reduced conditions typically found in subsurface environments, it is possible some of the reducing agents in the growth medium are responsible for this observed reduction despite the stable configuration of these metals commonly found in the subsurface.

The quantitative trends plotted on R_{CH4} -log(EC_{min}) in experiment 1 showed significant opposing trends in early-stage methanogenesis in Mm900 reactors when subjected to different environments; H₂-rich had higher R_{CH4} ratios in reactors with low-conductivity

while in H₂-limited reactors ratios were highest in reactors containing high-conductivity minerals. While it is not certain why these trends differ in these environments, the fact that only archaellated specimens showed significant trends in the early stages of methanogenesis may be important. Although this evidence was indirect, it suggested that minerals only make a difference when archaella are present during the early-stages methanogenesis. However, in experiment 2, pyrite was shown to enhance both k and $t_{1/2}$ with mutant reactors in H₂-rich conditions. Therefore, it is more likely that if electron shuttling is occurring it is happening in the cell structure with archaella facilitating attachment to mineral surfaces.

Although all of the kinetic parameters (c, k, $t_{1/2}$) plotted in experiment 2 had significant R²-values, only c provided consistent standard errors for the model. Since both k and $t_{1/2}$ are temporal parameters, it is likely that 4 days was too long of a duration for modeling the kinetics associated with these reactions. Higher resolution sampling at times below 1-day would likely improve the tightness of the error bars associated with the estimates for these parameters. The c values in pyrite reactors containing Mm900 were higher than limestone in both H₂-rich and H₂-limited environments as opposed to mutant strain $\Delta flaB2$ (limestone>pyrite in both environments). As opposed to the trends seen in experiment 1, c values for archaeallated strains were enhanced in the presence of pyrite in both environments. This would be expected if archaealla are directly involved in electron shuttling. However, in mutant reactors no trend would be expected with respect to respect to mineral conductivity instead of limestone outperforming pyrite. Therefore, from the evidence given it is likely that archaella are not used in direct electron shuttling, but as a means of motility and attachment to mineral surfaces.

Appendix A: Detailed Growth Media Preparation

Basal medium for formate growth (McF) of Methanococcus maripaludis

Modified from Feng Long, Whitman Lab, University of Georgia

1. Select appropriate glassware for experiment. For cultivation, 250-mL anaerobic bottles (Chemglass CLS-4217-03) were used to ensure enough headspace for 50 mL of inoculum.

2. Medium Composition:

Component	For 1 liter bottle	_
Glass-distilled water	500ml	
General salts solution	500ml	
K ₂ HPO ₄ , 14g/L	10ml	
Na acetate·3H ₂ O, 136 g/L	10ml	
Trace mineral solution	10ml	
Iron stock solution	5ml	
Rezasurin, 0.1g/100ml	1ml	
Sodium formate (NaCOOH)	27g	
Sodium bicarbonate (NaHCO ₃)	5.0g	

Combine medium ingredients in a 1-L flask and sparge with a stream of N_2 gas for 60 minutes.

3. Add 0.05g cysteine-HCl per 100ml and continue sparging until clear.

- 4. Transfer medium in anaerobic chamber, dispense 45 mL of medium into each culture vessel and seal tubes with blue butyl rubber stopper and aluminum crimp seal (VWR International# 98167-180, 30624-022, and 89167-184)
- 5. After removing from the anaerobic chamber, autoclave on gravity cycle for 25 minutes.
- 6. Prior to inoculation, flame sterilize and add 1 ml of 2.5% Na₂S·9H₂O (w/v) per 50 ml of medium.
- 7. Flame-sterilize once again and inoculate culture vessels with 5 mL of inoculum per 50 mL vessel.
- 8. Using a syringe-filter and an exhaust needle, cycle 80/20 v/v H₂/CO₂ through each reactor for 45 seconds and pressurize to 200 kPa.
- 9. Shake vigorously and place in 37°C incubator at least 5 days prior to use making sure to shake vigorously at least once each day.

Preparation of General Salts solution

Use 50 ml per 100 ml of medium (modified from Romesser et al., 1979)

Composition	g/L	Medium concentration (mM)
KCl	0.67	4.5
$MgCl_2 \cdot 6H_20$	5.50	13.5
$MgSO_4 \cdot 7H_20$	6.90	14.0
NH ₄ Cl	1.00	9.0
CaCl ₂ ·2H ₂ 0	0.28	0.95

Preparation of Iron Stock solution

Use 0.5 ml per 100 ml of medium

To a small screw top bottle, add 2.0 g of $Fe(NH_4)_2(SO_4)_2 \cdot 6H_20$. Then add 20 drops of concentrated HCl followed by 1000 ml of nanopure distilled water.

Preparation of Trace minerals solution

Use 1 ml per 1 liter of medium

Composition	g/L	Medium concentration(μM)
Nitriloacetic acid	1.5	78
$MnSO_4 \cdot 2H_20$	0.1	5.3
$Fe(NH_4)_2(SO_4)_2 \cdot H_20$	0.2	5.1
$CoCl_2 \cdot 6H_20$	0.1	4.2
$ZnSO_4 \cdot 7H_20$	0.1	3.5
CuSO ₄ ·5H ₂ 0	0.01	0.4
$NiCl_2 \cdot 6H_20$	0.025	1.1
Na_2SeO_3	0.2	11.6
$Na_2MoO_4 \cdot 2H_20$	0.1	4.1
$Na_2WO_4{\cdot}2H_20$	0.1	3.0

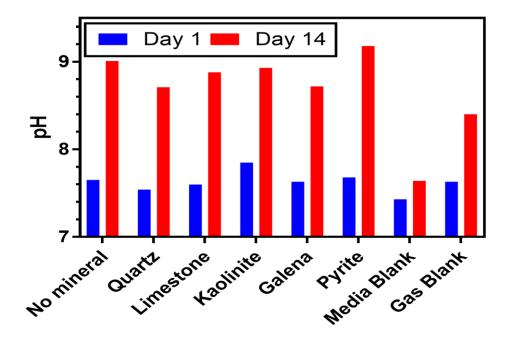
Neutralize the nitriloacetic acid to pH 6.5 with KOH, add minerals in order, allowing each one to dissolve before adding the next mineral, and adjust pH to 7.0.

Preparation of 2.5% Sodium Sulfide solution

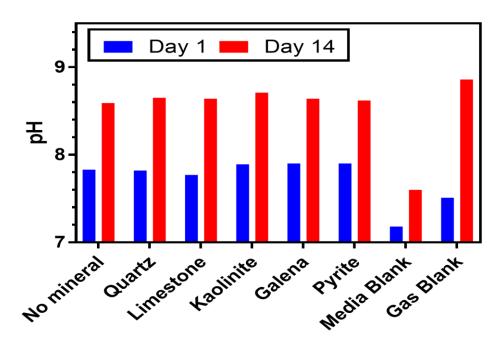
- (1) Add 100 ml nanopure H₂O to flask and mark the water line. To limit the formation of volatile hydrogen sulfide from sodium sulfide, add one pellet of NaOH. Add 10 ml more nanopure H₂O to flask.
- (2) Boil the 110ml nanopure H₂O while flushing flask with N₂ until the water level reaches the marked 100 ml water line.
- (3) Let flask cool while flushing with N₂, transfer the flask to the gassing station in the fume hood. Continue to flash with N₂.
- (4) While flask is cooling, weigh out slightly more than 2.5 g Na₂S·9H₂O. Wear gloves to do the subsequent steps in the fume hood. Clean the sodium sulfide crystal by briefly rinsing the crystal with nanopure H₂O followed by blotting dry with a paper towel. Reweigh the crystal to insure the final weight is 90-110% of the desired weight.
- (5) Add the cleaned and weighted sodium sulfide to the cooled flask while flushing with N₂ and mix until partially dissolved.
- (6) Stopper the flask, discontinue flushing, and dispense 5ml aliquots in 28 ml Balch tubes.
- (7) Seal and crimp tubes, clear the headspace with N₂ gas, and pressurize to 15 psi.
- (8) Autoclave on gravity cycle for 25 minutes.
- (9) Store these sodium sulfide tubes sealed in the fume hood; Discard if precipitant forms.

Appendix B: pH values for varying mineral types in early and late stages of growth

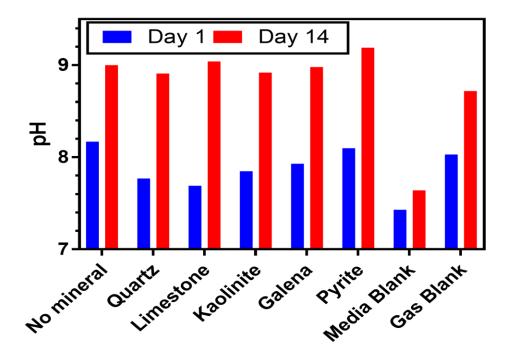
Mm900 in H2/CO2



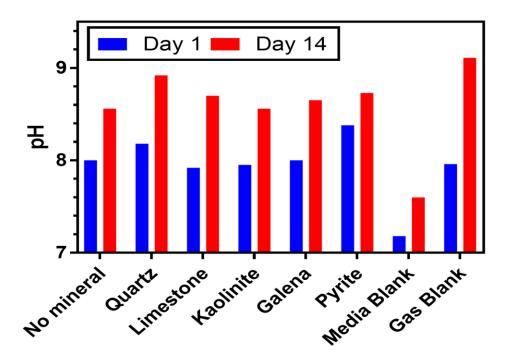
Mm900 in N2/CO2



△flaB2 in H2/CO2



∆flaB2 in N2/CO2



Appendix C: Overall CH₄ and CO₂ Concentrations on Day 14 (%)

H ₂ -rich conditions	CH ₄ (%)	CO ₂ (%)
<u>Mm900</u>		
Galena	1.55	0.68
Pyrite	2.90	0.29
Kaolinite	6.57	0.48
Quartz	5.85	0.94
Limestone	4.01	0.47
Non-mineral	4.91	1.20
∆flaB2		
Galena	2.38	0.29
Pyrite	8.81	0.18
Kaolinite	6.57	0.52
Quartz	6.47	0.49
Limestone	4.04	0.49
Non-mineral	13.2	0.27

H ₂ -limited conditions	CH ₄ (%)	CO ₂ (%)
<u>Mm900</u>		
Galena	22.7	1.97
Pyrite	19.4	1.16
Kaolinite	18.3	1.39
Quartz	20.9	1.97
Limestone	12.9	0.96
Non-mineral	16.8	1.95
∆flaB2		
Galena	14.1	0.95
Pyrite	15.3	1.04
Kaolinite	19.3	1.79
Quartz	9.23	0.12
Limestone	22.5	1.78
Non-mineral	18.9	1.65

REFERENCES

- (1) Zavala, J. A.; Gog, L.; Giacometti, R., Anthropogenic increase in carbon dioxide modifies plant–insect interactions. *Annals of Applied Biology* **2017**, *170*, (1), 68-77.
- (2) Yang, G. C.; Zhou, L.; Mbadinga, S. M.; Liu, J. F.; Yang, S. Z.; Gu, J. D.; Mu, B. Z., Formate-Dependent Microbial Conversion of CO2 and the Dominant Pathways of Methanogenesis in Production Water of High-temperature Oil Reservoirs Amended with Bicarbonate. *Frontiers in Microbiology* **2016**, *7*, 365.
- (3) Goldberg, D. S.; Takahashi, T.; Slagle, A. L., Carbon Dioxide Sequestration in Deep-Sea Basalt. *Proceedings of the National Academy of Sciences of the United States of America* **2008**, *105*, (29), 9920-9925.
- (4) Cuéllar-Franca, R. M.; Azapagic, A., Carbon capture, storage and utilisation technologies: A critical analysis and comparison of their life cycle environmental impacts. *Journal of CO2 Utilization* **2015**, *9*, (Supplement C), 82-102.
- (5) Koornneef, J.; Ramírez, A.; Turkenburg, W.; Faaij, A., The environmental impact and risk assessment of CO2 capture, transport and storage-an evaluation of the knowledge base using the DPSIR framework. *Energy Procedia* **2011**, *4*, (Supplement C), 2293-2300.
- (6) Bahrami, A.; Jamialahmadi, M.; Moghadasi, J.; Alimohammadi, N., Simulation study of CO2 sequestration potential of the Mary Lee coal zone, Black Warrior basin. *Environmental Earth Sciences* **2013**, *70*, (6), 2501-2509.
- (7) Muggeridge, A.; Cockin, A.; Webb, K.; Frampton, H.; Collins, I.; Moulds, T.; Salino, P., Recovery rates, enhanced oil recovery and technological limits. *Philosophical transactions*. *Series A, Mathematical, physical, and engineering sciences* **2014**, *372*.
- (8) Fang, Z.; Li, X.; Wang, G., A gas mixture enhanced coalbed methane recovery technology applied to underground coal mines. *Journal of Mining Science* **2013**, *49*, (1), 106-117.
- (9) Jin, C.; Liu, L.; Li, Y.; Zeng, R., Capacity assessment of CO2 storage in deep saline aquifers by mineral trapping and the implications for Songliao Basin, Northeast China. *Energy Science & Engineering* **2017**.
- (10) Celia, M. A., S. Bachu, J. M. Nordbotten, and K. W. Bandilla, Status of CO2 storage in deep saline aquifers with emphasis on modeling approaches and practical simulations. *Water Resour. Res* **2015**, *51*, 6846-6892.

- (11) Lau, M. C. Y.; Kieft, T. L.; Kuloyo, O.; Linage-Alvarez, B.; Van Heerden, E.; Lindsay, M. R.; Magnabosco, C.; Wang, W.; Wiggins, J. B.; Guo, L.; Perlman, D. H.; Kyin, S.; Shwe, H. H.; Harris, R. L.; Oh, Y.; Yi, M. J.; Purtschert, R.; Slater, G. F.; Ono, S.; Wei, S. W.; Li, L.; Lollar, B. S.; Onstott, T. C., An oligotrophic deep-subsurface community dependent on syntrophy is dominated by sulfur-driven autotrophic denitrifiers. *Proceedings of the National Academy of Sciences of the United States of America* **2016**, *113*, (49), E7927-E7936.
- (12) Kotelnikova, S., Microbial production and oxidation of methane in deep subsurface. *Earth Science Reviews* **2002**, *58*, (3), 367-395.
- (13) Harvey, O. R.; Qafoku, N. P.; Cantrell, K. J.; Wilkins, M. J.; Brown, C. F., Methanogenesis-induced pH-Eh shifts drives aqueous metal(loid) mobility in sulfide mineral systems under CO2 enriched conditions. *Geochimica et Cosmochimica Acta* **2016**, *173*, 232-245.
- (14) Goyal, N.; Padhiary, M.; Karimi, I. A.; Zhou, Z., Flux measurements and maintenance energy for carbon dioxide utilization by Methanococcus maripaludis. *Microbial Cell Factories* **2015**, *14*, 146.
- (15) Capone, D. G.; Reese, D. D.; Kiene, R. P., Effects of Metals on Methanogenesis, Sulfate Reduction, Carbon Dioxide Evolution, and Microbial Biomass in Anoxic Salt Marsh Sediments. *Applied and Environmental Microbiology* **1983**, *45*, (5), 1586-1591.
- (16) Lorowitz, W. H.; Nagle, D. P.; Tanner, R. S., Anaerobic oxidation of elemental metals coupled to methanogenesis by *Methanobacterium thermoautotrophicum*. *Environmental Science and Technology*. **1992**, *26*, (8), 1606-1610.
- (17) Conlette, O. C.; Emmanuel, N. E.; Chijoke, O. G., Methanogen Population of an Oil Production Skimmer Pit and the Effects of Environmental Factors and Substrate Availability on Methanogenesis and Corrosion Rates. *Microbial Ecology* **2016**, *72*, (1), 175-184.
- (18) Goyal, N.; Zhou, Z.; Karimi, I. A., Metabolic processes of Methanococcus maripaludis and potential applications. *Microbial Cell Factories* **2016**, *15*, (1), 107.
- (19) Jones, W. J.; Paynter, M. J. B.; Gupta, R., Characterization of Methanococcus maripaludis sp. nov., a new methanogen isolated from salt marsh sediment. *Archives of Microbiology* **1983**, *135*, (2), 91-97.
- (20) Jarrell, K. F.; Stark, M.; Nair, D. B.; Chong, J. P. J., Flagella and pili are both necessary for efficient attachment of Methanococcus maripaludis to surfaces. *FEMS Microbiology Letters* **2011**, *319*, (1), 44-50.
- (21) Lohner, S. T.; Deutzmann, J. S.; Logan, B. E.; Leigh, J.; Spormann, A. M., Hydrogenase-independent uptake and metabolism of electrons by the archaeon Methanococcus maripaludis. *ISME Journal* **2014**, *8*, (8), 1673-1681.

- (22) Sydow, A.; Krieg, T.; Mayer, F.; Schrader, J.; Holtmann, D., Electroactive bacteria—molecular mechanisms and genetic tools. *Applied Microbiology and Biotechnology* **2014**, *98*, (20), 8481-8495.
- (23) Bandara, W.; Kindaichi, T.; Satoh, H.; Sasakawa, M.; Nakahara, Y.; Takahashi, M.; Okabe, S., Anaerobic treatment of municipal wastewater at ambient temperature: Analysis of archaeal community structure and recovery of dissolved methane. *Water Research* **2012**, *46*, (17), 5756-5764.
- (24) Feng, Y. H.; Zhang, Y. B.; Quan, X.; Chen, S., Enhanced anaerobic digestion of waste activated sludge digestion by the addition of zero valent iron. *Water Research* **2014**, *52*, 242-250.
- (25) Takeshita, T., A cost-optimal scenario of CO2 sequestration in a carbon-constrained world through to 2050. *Natural Science* **2013**, *5*, (2), 313-319.
- (26) Kato, S.; Hashimoto, K.; Watanabe, K., Methanogenesis facilitated by electric syntrophy via (semi)conductive iron-oxide minerals: Methanogenesis via (semi)conductive minerals. *Environmental Microbiology* **2012**, *14*, (7), 1646-1654.
- (27) Zhou, S.; Xu, J.; Yang, G.; Zhuang, L., Methanogenesis affected by the co-occurrence of iron(III) oxides and humic substances. *FEMS Microbiology Ecology* **2014**, *88*, (1), 107-120.
- (28) Jiang, S. H.; Park, S.; Yoon, Y.; Lee, J. H.; Wu, W. M.; Dan, N. P.; Sadowsky, M. J.; Hur, H. G., Methanogenesis Facilitated by Geobiochemical Iron Cycle in a Novel Syntrophic Methanogenic Microbial Community. *Environmental Science and Technology* **2013**, *47*, (17), 10078-10084.
- (29) Malvankar, N. S.; King, G. M.; Lovley, D. R., Centimeter-long electron transport in marine sediments via conductive minerals. *ISME Journal* **2015**, *9*, (2), 527-531.
- (30) Qiao, S.; Tian, T.; Qi, B.; Zhou, J., Methanogenesis from wastewater stimulated by addition of elemental manganese. *Scientific Reports* **2015**, *5*, 12732.
- (31) Palacky, G. J., Resistivity Characteristics of Geologic Targets. *Electromagnetic Methods in Applied Geophysics* **1983**, 52-129.
- (32) Pridmore, D. F. a. S. R. T., The electrical resistivity of galena, pyrite, and chalcopyrite. *American Mineralogist* **1976**, *61*, 248-259.
- (33) Belay, N.; Daniels, L., Elemental metals as electron sources for biological methane formation from CO₂. *Antonie van Leeuwenhoek International Journal of General and Moloecular Microbiology.* **1990,** *57*, (1), 1-7.

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ABSTRACT

CO₂ TO NATURAL GAS (CH₄) CONVERSION IN METHANOGEN-MINERAL SYSTEMS: THE ROLE OF MINERAL TYPE

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Although the mechanisms are not completely understood, the need for both archaella and pili for attachment of *Methanoccocus maripaludis* to mineral surfaces has been well established. In addition, the uses for these pili and/or archaella are still not completely understood, but it has been proposed that they could function to preferentially attach to conductive surfaces allowing for electron shuttling. In reactors containing minerals of varying mineral electrical conductivity -pyrite >galena >kaolinite >quartz>limestone-wildtype *M. maripaludis* (Mm900) and non-archaellated mutant ($\Delta flaB2$) were examined in H₂-rich and H₂-limited environments to determine trends in early-stage methanogenesis. Building on these trends, a kinetic model of these early-stage trend was created based on the fitted parameters (c, k, and $t_{1/2}$).